Risk assessment of leukaemia and occupational exposure to benzene

G M H SWAEN, J M M MEIJERS

From the Department of Occupational Health and Toxicology, University of Limburg, 6200 MD Maastricht, The Netherlands

ABSTRACT  Experimental toxicological studies have offered clear evidence that benzene induces haematopoietic neoplasms, and it is generally accepted that exposure to benzene is a risk factor for leukaemia, in particular for acute non-lymphatic leukaemia. Quantitative aspects of benzene risk assessment are still a matter of controversy, however. In several risk assessments an estimated 50 deaths from leukaemia per 1000 deaths would arise from exposures to benzene of 10 ppm during the working life of 30 years. The assessment presented in this paper leads to lower estimates, which are in agreement with the weak toxicological data. Furthermore, an approach is presented to incorporate the results of low exposure epidemiological studies into the process of quantitative risk assessment.

In the past decades a substantial body of knowledge has accumulated regarding the long term health effects of occupational exposure to benzene. In addition to experimental data, this consists of epidemiological data mainly from retrospective cohort studies of workers occupationally exposed to benzene. Except for the ability of benzene to cause non-malignant blood anomalies, the main chronic health effect of concern is the increased mortality due to leukaemia and, in particular, acute non-lymphatic leukaemia: several investigators have stated that this is the only increased risk after exposure to benzene. Nevertheless, other investigators have pointed out that benzene induces a variety of malignant neoplasms, including lymphatic leukaemia and lymphomas.1-3 The relation between exposure to high concentrations of benzene and the risk of leukaemia is well documented and generally accepted. Case reports and epidemiological confirmations have led to regulatory legislation throughout the world in the field of occupational exposure, and exposures experienced by the general population.4-10 Several risk assessments have been conducted for occupational exposures.11-15 In a recent review the authors presented a risk assessment based on the results of a retrospective cohort study of 1006 exposed workers, applying a linear dose-response relation.15

Following a request from the Minister of Welfare, Public Health and Culture, a committee of the Health Council of The Netherlands has drawn up a somewhat different risk assessment for the risk of leukaemia and exposure to benzene in ambient air. It is the purpose of this article to apply this risk assessment to occupational exposures to benzene.

Several weaknesses of epidemiological data can be identified that may form a source of controversy. The first weakness is the lack of accurate and reliable data regarding the exposure levels that existed in the past and have been experienced by the studied cohorts. This is not a particular problem in the course of a qualitative risk assessment: in other words to assess exposure to high concentrations of a certain chemical is a risk factor for a particular neoplasm. In the course of a quantitative risk assessment, however, the estimation of past exposure will have substantial impact on the outcome of the assessment. A second weakness of epidemiological data is the lack of information regarding the mechanisms, leading to the increased risk observed. Insight into the mechanism that is at work may be of great importance in the determination of the extrapolation model to be applied to assess the risks related to low exposure levels. The choice of the extrapolation model may also have a great impact on the outcome of the risk assessment. A major issue is whether or not the chemical should be considered a genotoxic or not. In the case of a non-genotoxic agent, it is customary to take a no-effect level approach and...
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apply a safety factor to be on the safe side. In the case of a genotoxic agent a non-threshold approach is generally applied because it may be argued that even a small dose may provide a carcinogenic effect. The non-threshold approach does not imply that the extrapolation model is linear or logistic. After a brief review of experimental toxicologic data of the carcinogenic potency of benzene, several remarks will be made regarding the available epidemiological data. Then the risk assessment proposed by the Health Council of The Netherlands will be described.

Experimental toxicological data

Radioactive labelled benzene has been reported to be incorporated into nucleic acids in the liver, spleen, bone marrow, and kidney cell of mice and rats. Several recent publications have made covalent binding of benzene metabolites to DNA even more plausible. These investigations were conducted by means of isolated bone marrow mitochondria and their relevance for in vivo genotoxicity is at least questionable. Investigations of the effects of benzene on DNA repair and the occurrence of unscheduled DNA-synthesis in mammalian cell cultures and in bacteria have so far not yielded any evidence of DNA interaction. Despite extensive efforts there are no indications that benzene is capable of inducing gene mutations in bacteria, yeasts, drosophila, or mammalian cells if tests not capable of discriminating between gene mutations and recombinations are excluded. Benzene has been shown to be capable of inducing chromosome aberrations in yeast, fungi, insects, and somatic mammalian cells, including human lymphocytes. Chromosome breakage phenomena such as translocations have been found in vitro and in vivo in mammalian cells. These chromosome aberrations were observed with benzene concentrations lower than 10 ppm. Several in vivo experiments indicate that the sensitivity of laboratory animals to clastogenic effects indicate that the sensitivity of laboratory animals to clastogenic effects decreases with containing exposure. Benzene has been tested for neoplastic transformation in a variety of test systems. Only in Syrian hamster cells was an increase in the incidence of transformed cell clones observed. Benzene is also capable of increasing the occurrence of tumours in Zymbal’s gland, a comparatively rare organ for tumour formation in laboratory animals. On the basis of the multiple site carcinogenicity in animals, benzene may be considered as a potent carcinogen in animals and the carcinogenic process should in principle be viewed as a stochastic process.

Epidemiological data

After case reports were published indicating a possible relation between exposure to benzene and the occurrence of leukaemia, many epidemiological studies in this field were started and an extensive review has recently been published. The findings of these studies leave no doubt that workers exposed in the past to high concentrations of benzene have experienced an increased mortality from leukaemia. Both case-control studies and retrospective cohort studies have confirmed this relation. Several studies, however, in low exposure groups have reported contradictory results. For instance, in a retrospective cohort study of about 13 500 workers in the rubber industry an excess mortality risk for lymphatic leukaemia was observed and not for non-lymphatic leukaemia as has been reported by other epidemiologists. In several other large studies of workers exposed to low concentrations of benzene no indications were found that these cohorts had experienced increased mortality for leukaemia. In a study of 38 800 workers employed in the petrochemical industry 18 cases of leukaemia were observed compared with an expected number of 23. In a study of similar size (34 781 workers) in the oil industry 30 deaths from leukaemia were observed, compared with an expected number of 32. In this study, however, a dose response relation appeared to exist, which was shown by means of a nested case-control study design. Again this case-control study confirmed the existence of a risk of leukaemia after exposure to high concentrations of benzene. A third study of workers exposed to benzene that should also be regarded as negative is the study conducted by Parkes et al in the British rubber industry. In this study of 33 815 workers, 31 cases of leukaemia were observed compared with an expected number of 28.

As has been pointed out by Hernberg, the interpretation of negative results of epidemiological studies is complicated. With a negative outcome, it is not always clear whether the finding is an effect of methodological deficiencies or if there actually is no increased risk present in the cohort and exposure level under investigation.

The greatest weaknesses of the three large studies of low exposures are the lack of quantitative exposure data and the possibility that a proportion of the “exposed cohort” had not been exposed to benzene at all. Despite these weaknesses, the committee of the Health Council of The Netherlands decided not to disregard these findings in establishing a risk assessment. Thus the committee was left with the task of incorporating positive results of high exposure studies and negative results of low exposure studies in a risk assessment, low exposures being defined as lower than
a time weighted average exposure of 10 ppm which is the current threshold unit value in many countries. This contradictory conception has also been supported by the European Chemical Industry Ecology and Toxicology Centre.35 As a consequence it was decided not to estimate an overall relative risk for leukaemia after exposure to benzene, as has been done in the evaluation of vinyl chloride.36 Two separate risk assessments were made, both based on the findings of epidemiological studies. The first departed from the findings in high exposure studies, the second from the findings in low exposure studies. Although the risk assessment conducted by the committee was intended for risks experienced by the general population, it may also be applied to the occupational environment.

RISK ASSESSMENT BASED ON STUDIES OF HIGHLY EXPOSED WORKERS

Before embarking on a risk assessment several assumptions must be made. The first deals with the magnitude of risk after a particular exposure dose. The second is concerned with the extrapolation model. The first assumption may be derived from several cohort studies. A reasonable estimate of this risk, based on Rinsky's risk assessment, is that workers exposed on average to 40 ppm over a period of 10 years have experienced a fivefold increase in the risk of dying from acute non-lymphatic leukaemia (SMR = 500).13 Assuming a stochastic working mechanism, if a doubling of the dose also implies a doubling of the risk, a linear extrapolation model is suitable. Such a model may be formulated as:

$$\text{SMR} = 100 + b \times d$$

where SMR is the standardised mortality ratio (observed/expected × 100), b is the tangent of the angle between the straight line and the horizontal axis (dose), and d is the benzene dose in ppm-years. By means of this model the number of additional deaths from leukaemia per 1000 deaths which may arise from a benzene dose of 300 ppm-years may be calculated, as done by Austin et al35:

$$500 = 100 + b \times 400 \rightarrow b = 1$$

if d = 300 ppm-years then SMR =

$$100 + 1 \times 300 = 400$$

Thus workers having received a total benzene dose of 300 ppm-years may experience a SMR of 400.

In a western country such as the Netherlands the age adjusted death rate for acute non-lymphatic leukaemia in men is 1.6 per 100 00037 and the total mortality is 923.2 per 100 000. This implies that about 1.7 deaths per 1000 deaths are due to acute non-lymphatic leukaemia. In a population in which an SMR of 400 exists 4 × 1.7 = 6.8 deaths due to acute non-lymphatic leukaemia may occur per 1000 deaths, which is an excess of about five deaths. This estimate differs from that proposed by Austin35 which is in the range of about 50 excess deaths per 1000 deaths.

RISK ASSESSMENT BASED ON THE RESULTS OF LOW EXPOSURE STUDIES

There are three large epidemiological studies of workers exposed mainly to low concentrations of benzene.30-33 Thorpe conducted a study of 38 000 workers in petroleum refineries.30 Although this study has several methodological weaknesses, the study could have detected a risk if there was one. Rushlow and Alderson studied the mortality patterns of 35 000 male employees with a minimum of one year's continuous service in the United Kingdom oil industry.31 There was no excess of leukaemia. A third large study of workers potentially exposed to benzene focused on the rubber industry. Parkes et al followed up 34 600 British rubber workers for whom no increased risks could be detected.33 In these three studies combined a total number of 18 + 30 + 31 = 79 deaths from leukaemia were observed. A 95% two sided confidence interval may be calculated around this combined finding, which is:

$$\frac{\ln(\text{SMR}) \pm 1.96}{1/\text{obs}} = \frac{(4.055 \pm 0.1125)}{92} = 89 + 152$$

This upper limit can serve as a point of departure for the risk assessment. Again no accurate data on past exposures to benzene are available. Since 10 ppm was the threshold limit value in the early 1970s, it may be reasonable to use 5 ppm during a period of ten years as an estimate of the exposure. Subsequently it is possible to calculate b in the linear model:

$$\text{SMR} = 100 + b \times d$$

$$112 = 100 + b \times 50$$

$$b = 12/50 = 0.24$$

which may be used to calculate the SMR given a total dose of 300 ppm-years.

$$\text{SMR} = 100 + 0.24 \times 300$$

$$\text{SMR} = 172$$

Since the confidence limits were based on total leukaemia mortality, it seems appropriate to extrapolate for total leukaemia mortality, even though the risk is probably restricted to acute non-lymphatic leukaemia. The age adjusted death rate for total leukaemia among men in The Netherlands is 7.6 per 100 000. The total mortality is 923.2 per 100 000, which implies that 8.2 deaths per 1000 deaths will be due to leukaemia if no exposure to benzene occurs. Thus in a group of workers experiencing a dose of 300 ppm-years of benzene, associated with an estimated SMR of 172, 8.7 × 1 - 1.72 = 15 deaths from leukaemia may occur, of which 8.7 are attributable to the "natural" background incidence of leukaemia.

By analogy, a risk assessment may be made limiting...
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the potential risk to acute non-lymphatic leukaemia. This perspective is probably more realistic given the observations cited earlier in this article.

In The Netherlands the annual age adjusted death rate from acute non-lymphatic leukaemia is 1.6 per 100 000 men. For total leukaemia this rate is 7.6 per 100 000 men, which implies that approximately 21% of deaths from leukaemia occurring in men are of the acute non-lymphatic type. Thus generalising this figure to the three low exposure studies, an estimated 17 (21%) of the 79 deaths from leukaemia observed in these studies may have been of the acute non-lymphatic type giving a two sided 95% confidence interval of 62–161 for the SMR. This upper limit may again serve as the SMR resulting from a benzene dose of 50 ppm-years. Next b may be calculated:

\[ \text{SMR} = \frac{100 + b \times d}{161 + 100 + b \times 50} = b = 61/50 = 1.22 \]

If a dose of 300 ppm-years is experienced an SMR for acute non-lymphatic leukaemia may be estimated by means of the linear model as follows:

\[ \text{SMR} = \frac{100 + 1.21 \times 300}{466} \]

This implies that instead of the 1-7 deaths from acute non-lymphatic leukaemia an estimated eight deaths may occur per 1000 deaths, of which 1-7 are attributable to the "natural" background incidence of acute non-lymphatic leukaemia.

Discussion

There is no doubt that relatively long exposure to high concentrations of benzene increases the risk of dying from acute non-lymphatic leukaemia. This observation has frequently been confirmed in epidemiological studies. In studies of workers exposed to low concentrations, however, this relation has been absent or was confined to those workers who had been exposed to high concentrations in the past. It did not seem justified to disregard these negative findings in risk assessments as being merely an effect of methodological shortcomings in these studies. Several risk assessments have been conducted, giving risks in the range of 50 excess deaths from leukaemia per 1000 deaths among workers exposed to 300 ppm-years of benzene. The risk assessment presented in this article clearly leads to lower results than, for instance, the one prepared by Austin et al. Both assessments are based on several assumptions which are difficult, if not impossible, to verify. Thus it remains of great importance to conduct updates of cohorts already identified as having been exposed to benzene. Although it is up to the policy makers to decide whether or not to alter the current occupational standard, perhaps a guideline might be given as to what level of risk would be regarded as acceptable in an occupational setting. This is more a matter of ethics than of science. Nevertheless, it should be remembered that benzene is not the only chemical to which workers can be exposed, and that the interindividual susceptibility for the haematopoietic effects of benzene can differ widely. Perhaps one additional death per 1000 deaths after a working life of exposure can serve as a guideline.

Requests for reprints to: G M H Swaen, Department of Occupational and Environmental Health and Toxicology, University of Limburg, PO Box 616, 6200 MD Maastricht, The Netherlands.

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