Occupational exposures and pancreatic cancer: a meta-analysis

I Anneli Ojajärvi, Timo J Partanen, Anders Ahlbom, Paolo Boffetta, Timo Hakulinen, Nadia Jourenkova, Timo P Kauppinen, Manolis Kogevinas, Miquel Porta, Harri U Vainio, Elisabete Weiderpass, Catharina H Wesseling

Abstract

Objectives—Consolidation of epidemiological data on pancreatic cancer and workplace exposures.

Methods—Publications during 1969–98 were surveyed. Studies without verified exposures were excluded. Meta-analyses were conducted on data from 92 studies covering 161 populations, with results for 23 agents or groups of agents. With a standard format, five epidemiologists extracted risk estimates and variables of the structure and quality of each study. The extracted data were centrally checked. Random meta-models were applied.

Results—Based on 20 populations, exposure to chlorinated hydrocarbon (CHC) solvents and related compounds was associated with a meta-risk ratio (MRR) of 1.4 (95% confidence interval (95% CI) 1.0 to 1.8). Nickel and nickel compounds were considered in four populations (1.9; 1.2 to 3.2). Excesses were found also for chromium and chromium compounds (1.4; 0.9 to 2.3), polycyclic aromatic hydrocarbons (PAHs) (1.5; 0.9 to 2.5), organochlorine insecticides (1.5; 0.6 to 3.7), silica dust (1.4; 0.9 to 2.0), and aliphatic and alicyclic hydrocarbon solvents (1.3; 0.8 to 2.8). Evidence on pancreatic carcinogenicity was weak or non-positive for the following agents: acrylonitrile (1.1; 0.6 to 2.2); arsenic (1.0; 0.6 to 1.5); asbestos (1.1; 0.9 to 1.5); diesel engine exhaust (1.0; 0.9 to 1.3); electromagnetic fields (1.1; 0.8 to 1.4); formaldehyde (0.8; 0.5 to 1.0); flour dust (1.1; 0.3 to 3.2); cadmium and cadmium compounds (0.7; 0.4 to 1.4); gasoline (1.0; 0.8 to 1.2); herbicides (1.0; 0.8 to 1.3); iron and iron compounds (1.3; 0.7 to 2.5); lead and lead compounds (1.1; 0.8 to 1.5); man-made vitreous fibres (1.0; 0.6 to 1.0); oil mist (0.9; 0.8 to 1.0); and wood dust (1.1; 0.9 to 2.5). The occupational aetiological fraction of pancreatic cancer was estimated at 12%. In a subpopulation exposed to CHC solvents and related compounds, it was 29%; to chromium and chromium compounds, 23%; to nickel and nickel compounds, 47%; to insecticides, 33%; and to PAHs, 33%.

Conclusion—Occupational exposures may increase risk of pancreatic cancer. High quality studies are called for on interactions between occupational, environmental, and lifestyle factors as well as interactions between genes and the environment.

Keywords: pancreatic cancer; occupational exposure; meta-analysis

Some 180 000 pancreatic cancers are registered annually in the world. It is highly and rapidly fatal and represents the fifth leading cause of deaths from cancer in industrialised countries and is 50%–100% more common in men than in women. It is not consistently associated with socioeconomic status within national populations, although there is a tendency toward higher age adjusted risk in richer than poorer countries. Incidence has risen in industrialised countries since the 1960s and subsequently levelled off in several populations.

The causes of pancreatic cancer are mostly unknown. Tobacco smoking is the single established common cause. The proportion of cases attributable to smoking has been estimated at 5%–50%, depending on the population. Epidemiology of pancreatic cancer has suffered from bias due to high misclassification rate, notably prominent when case definition has been based only on death certificates. This has resulted in inconsistencies in the results on the aetiological role of environmental and occupational determinants of pancreatic cancer.

We identified published epidemiological studies about pancreatic cancer and job titles, industrial branches, and occupational exposures, and conducted a meta-analysis of the role of 23 chemical or physical agents present in the working environment that affect the aetiology of pancreatic cancer.

Materials and methods

The literature search covered the Medline, Toxline, and Cancerlit databases for the period 1969 to May 1998, with the following search conditions:

1. (occupational OR agriculture) AND neoplasms AND morbidity
2. (occupational OR agriculture) AND neoplasms AND mortality NOT morbidity
3. (occupational OR agriculture) AND neoplasms AND incidence NOT mortality
4. (pancreatic OR digestive) AND occupational
5. (pancreatic OR digestive) AND case AND (control OR referent)
The search was accompanied by a scan of the lists of reference of the identified studies. In all, 1902 studies were identified. A total of 365 studies remained after exclusion of studies that did not report on pancreatic cancer; that did not report the most recent update; that reported insufficient data for the meta-analysis; that did not report data for any job or occupational agent; that did not report original results (reviews); that reported on part of a larger population reported elsewhere; and that reported on job categories or agent categories too broad or outside our list of job titles and agents. The agents were based on the FINJEM job exposure matrix. The list of job titles covered 150 entries in the Finnish social status categories 3, 4, and 5. Data for categories 1 and 2 represented the highest social categories and were excluded because the relevant occupational and physical exposures were minimal or non-existent. The chemical and physical agents considered were the following 23 agents or groups of agents: aliphatic and alicyclic hydrocarbon solvents; aromatic hydrocarbon solvents (excluding aromatic amines); arsenic; asbestos; cadmium and cadmium compounds; chlorinated hydrocarbon (CHC) solvents and related compounds (excluding organochlorine insecticides); chromium and chromium compounds; diesel engine exhaust; electromagnetic fields, flour dust, formaldehyde, fungicides, gasoline, herbicides, insecticides, iron and iron compounds; lead and lead compounds; man made vitreous fibres; nickel compounds; oil mist (including machining fluid and cutting fluid); polycyclic aromatic hydrocarbons (PAH); silica dust; and wood dust.

The studies were divided into (a) agent specific studies with direct risk estimates for one or several of the 23 agents, or for job titles with verified exposure(s) to the agent(s), and (b) job specific studies without risk estimates for any of the selected agents but instead for one or more of the 150 job categories without verified exposure(s) to the agent(s). This report is based on the data set from agent specific studies only (92 studies presenting data for 161 di

<table>
<thead>
<tr>
<th>Table 1</th>
<th>Characterisation of the 92 agent specific studies</th>
</tr>
</thead>
<tbody>
<tr>
<td>Study type:</td>
<td></td>
</tr>
<tr>
<td>Administrative (linkage of administrative records or PMR/PCMR/MOR studies)</td>
<td>23</td>
</tr>
<tr>
<td>Industrial cohort</td>
<td>88</td>
</tr>
<tr>
<td>Industry based (nested) case-control study</td>
<td>7</td>
</tr>
<tr>
<td>Population or hospital based case-control study</td>
<td>43</td>
</tr>
<tr>
<td>Cases:</td>
<td></td>
</tr>
<tr>
<td>Exocrine pancreatic cancers only</td>
<td>32</td>
</tr>
<tr>
<td>All pancreatic cancers</td>
<td>127</td>
</tr>
<tr>
<td>Unspecified</td>
<td>2</td>
</tr>
<tr>
<td>Diagnosis of cases:</td>
<td></td>
</tr>
<tr>
<td>Histological</td>
<td>47</td>
</tr>
<tr>
<td>Other (clinical: radiology, necropsy, etc)</td>
<td>2</td>
</tr>
<tr>
<td>Mortality files</td>
<td>96</td>
</tr>
<tr>
<td>Mixed</td>
<td>9</td>
</tr>
<tr>
<td>Unknown</td>
<td>7</td>
</tr>
<tr>
<td>Ascertainment of cases:</td>
<td></td>
</tr>
<tr>
<td>Mortality files</td>
<td>98</td>
</tr>
<tr>
<td>Cancer registry files</td>
<td>40</td>
</tr>
<tr>
<td>Hospital records</td>
<td>21</td>
</tr>
<tr>
<td>Mixed</td>
<td>1</td>
</tr>
<tr>
<td>Unspecified</td>
<td>1</td>
</tr>
<tr>
<td>Sex:</td>
<td></td>
</tr>
<tr>
<td>Men</td>
<td>112</td>
</tr>
<tr>
<td>Women</td>
<td>8</td>
</tr>
<tr>
<td>Both men and women or unspecified</td>
<td>41</td>
</tr>
<tr>
<td>Risk measure:</td>
<td></td>
</tr>
<tr>
<td>SMR (standardised mortality ratio)</td>
<td>68</td>
</tr>
<tr>
<td>SIR (standardised incidence ratio)</td>
<td>17</td>
</tr>
<tr>
<td>PMR (proportional mortality ratio)</td>
<td>18</td>
</tr>
<tr>
<td>MOR (mortality odds ratio)</td>
<td>1</td>
</tr>
<tr>
<td>HR (hazard ratio)</td>
<td>3</td>
</tr>
<tr>
<td>OR (odds ratio)</td>
<td>50</td>
</tr>
<tr>
<td>RR (risk ratio)</td>
<td>4</td>
</tr>
<tr>
<td>Source of exposure data:</td>
<td></td>
</tr>
<tr>
<td>Industrial hygiene measurements</td>
<td>4</td>
</tr>
<tr>
<td>Job exposure matrix</td>
<td>15</td>
</tr>
<tr>
<td>Expert assessment</td>
<td>25</td>
</tr>
<tr>
<td>Job title†</td>
<td>57</td>
</tr>
<tr>
<td>Other‡</td>
<td>19</td>
</tr>
<tr>
<td>Mixed</td>
<td>37</td>
</tr>
<tr>
<td>Unknown</td>
<td>4</td>
</tr>
<tr>
<td>Time reference of exposure:</td>
<td></td>
</tr>
<tr>
<td>Last or around diagnosis</td>
<td>17</td>
</tr>
<tr>
<td>Earlier cross section</td>
<td>9</td>
</tr>
<tr>
<td>Lifetime longitudinal</td>
<td>47</td>
</tr>
<tr>
<td>Less than lifetime longitudinal</td>
<td>80</td>
</tr>
<tr>
<td>Other</td>
<td>1</td>
</tr>
<tr>
<td>Unknown</td>
<td>7</td>
</tr>
<tr>
<td>Total</td>
<td>161</td>
</tr>
</tbody>
</table>

*Ratio of risks or cumulative incidences in exposed and unexposed cohorts.  †Job titles with specifically verified exposure(s) to the relevant agent(s).  ‡Employment duration, biological monitoring, employer record, registry of chemicals, self reporting.
Population aetiological fractions were estimated as $\text{PEF} = \frac{p-p}{\text{MRR}}$, and aetiological fractions among the exposed groups as $\text{EEF} = 1 - \frac{1}{\text{MRR}}$ for each agent where excesses were observed. The proportion of pancreatic cancer cases who were exposed ($p$), was calculated as $\sum$ exposed cases/$\sum$ all cases, summing over studies that provided the necessary numbers. The PEFs were based on MRRs from population based and hospital based case-control studies only, whereas the EEFs were based on MRRs from all except proportional studies. Aetiological fractions were not calculated for agents with no indication of increased risk.

The data were organised and analysed by populations rather than studies. This was because most studies considered more than one subpopulation defined by exposure. A total of 161 populations were covered in agent specific studies (table 1). Most (85) populations were from North America, closely followed by Western Europe (63). There were few populations from Central and Eastern Europe, Oceania, and none from Middle and South America, Asia, or Africa. The annual number of studies has been rising considerably during 1969–8 (two studies during 1969–79; 58 studies during 1980–9, and 101 studies during 1990–8).

Industrial cohort (88 populations) and case-control (50 populations) studies were the most common agent specific studies (table 1). The cases represented predominantly all pancreatic cancers, irrespective of type. As most studies considered mortality, the diagnosis was in most populations obtained from the death record. Less than 5% of the populations were women. Exposures were assessed in 57 populations through job titles; in 25 through expert assessments; in 15 through job exposure matrices (JEMs); and in 60 through other, mixed, or unexplained methods. Industrial hygiene measurements were explicitly applied in four populations only. Exposure assessment was longitudinal for 127 populations.

Reference populations in cohort studies represented predominantly national or other large populations (57 studies). Fewer studies (23) used local populations as the reference; even fewer (seven) used an internal reference. Follow up for case ascertainment began usually during 1940–79 (90 studies) and spanned in most studies a period of $\geq$10 years (98 studies). Losses to follow up in agent specific cohort studies were minor or moderate (<5% in 44 studies but unknown in 51 studies) but were more marked in job branch cohort studies. Agent specific cohorts were rather evenly distributed between entry cohorts, cross sectional cohorts, and mixed cohorts.

Case-control studies used variable periods of case ascertainment. Most agent specific case-control studies (44) used cancer or population controls. Response rates were $\leq$80% for the cases, and $\leq$90% in the controls. Sixteen studies did not report response rates.

### Table 2 Populations (n), metarisk estimates (MRRs) (95% CIs), ranges of point estimates and p values for heterogeneity by occupational agents (proportional studies excluded: simple random effects models with no covariates)

<table>
<thead>
<tr>
<th>Agent</th>
<th>n</th>
<th>MRR</th>
<th>95% CI</th>
<th>Range of point estimate</th>
<th>p Value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Aliphatic and alicyclic hydrocarbon solvents</td>
<td>2</td>
<td>1.3</td>
<td>0.8 to 2.0</td>
<td>1.0–1.6</td>
<td>0.09</td>
</tr>
<tr>
<td>Aromatic hydrocarbon solvents</td>
<td>13</td>
<td>1.0</td>
<td>0.8 to 1.3</td>
<td>0.5–2.9</td>
<td>0.2</td>
</tr>
<tr>
<td>Arsenic</td>
<td>4</td>
<td>1.0</td>
<td>0.6 to 1.6</td>
<td>0.9–1.4</td>
<td>0.9</td>
</tr>
<tr>
<td>Asbestos</td>
<td>24</td>
<td>1.1</td>
<td>0.9 to 1.4</td>
<td>0.5–3.6</td>
<td>0.004</td>
</tr>
<tr>
<td>Cadmium and cadmium compounds</td>
<td>2</td>
<td>0.7</td>
<td>0.4 to 1.4</td>
<td>0.7–0.8</td>
<td>0.9</td>
</tr>
<tr>
<td>Chlorinated hydrocarbon solvents and related compounds</td>
<td>20</td>
<td>1.4</td>
<td>1.0 to 1.8</td>
<td>0.3–4.9</td>
<td>0.05</td>
</tr>
<tr>
<td>Chromium and chromium compounds</td>
<td>9</td>
<td>1.4</td>
<td>0.9 to 2.3</td>
<td>0.6–20</td>
<td>0.2</td>
</tr>
<tr>
<td>Diesel engine exhaust</td>
<td>7</td>
<td>1.0</td>
<td>0.9 to 1.2</td>
<td>0.5–1.4</td>
<td>0.2</td>
</tr>
<tr>
<td>Electromagnetic fields</td>
<td>5</td>
<td>1.1</td>
<td>0.8 to 1.4</td>
<td>0.6–2.4</td>
<td>0.04</td>
</tr>
<tr>
<td>Flour dust</td>
<td>1</td>
<td>1.1</td>
<td>0.3 to 3.2</td>
<td>0.05</td>
<td></td>
</tr>
<tr>
<td>Formaldehyde</td>
<td>5</td>
<td>0.8</td>
<td>0.5 to 1.0</td>
<td>0.5–1.0</td>
<td>0.3</td>
</tr>
<tr>
<td>Fungicides</td>
<td>2</td>
<td>1.3</td>
<td>0.4 to 3.8</td>
<td>1.3–1.4</td>
<td>0.9</td>
</tr>
<tr>
<td>Gasoline</td>
<td>4</td>
<td>1.0</td>
<td>0.8 to 1.2</td>
<td>0.2–1.1</td>
<td>0.5</td>
</tr>
<tr>
<td>Herbicides</td>
<td>10</td>
<td>1.0</td>
<td>0.8 to 1.3</td>
<td>0.6–5.9</td>
<td>0.3</td>
</tr>
<tr>
<td>Insecticides</td>
<td>3</td>
<td>1.5</td>
<td>0.6 to 3.7</td>
<td>0.8–21.0</td>
<td>0.1</td>
</tr>
<tr>
<td>Iron and iron compounds</td>
<td>1</td>
<td>1.3</td>
<td>0.7 to 2.5</td>
<td>0.05</td>
<td></td>
</tr>
<tr>
<td>Lead and lead compounds</td>
<td>4</td>
<td>1.1</td>
<td>0.8 to 1.5</td>
<td>1.0–1.4</td>
<td>1.0</td>
</tr>
<tr>
<td>Man made vitreous fibres</td>
<td>5</td>
<td>1.0</td>
<td>0.6 to 1.6</td>
<td>0.3–6.8</td>
<td>0.03</td>
</tr>
<tr>
<td>Nickel and nickel compounds</td>
<td>4</td>
<td>1.9</td>
<td>1.2 to 3.2</td>
<td>1.2–3.6</td>
<td>0.9</td>
</tr>
<tr>
<td>Oil mist</td>
<td>6</td>
<td>0.9</td>
<td>0.8 to 1.0</td>
<td>0.3–1.0</td>
<td>0.5</td>
</tr>
<tr>
<td>Polycyclic aromatic hydrocarbons (PAHs)</td>
<td>4</td>
<td>1.5</td>
<td>0.9 to 2.9</td>
<td>1.3–3.6</td>
<td>0.8</td>
</tr>
<tr>
<td>Silica dust</td>
<td>3</td>
<td>1.4</td>
<td>0.9 to 2.0</td>
<td>1.1–2.0</td>
<td>0.2</td>
</tr>
<tr>
<td>Wood dust</td>
<td>4</td>
<td>1.1</td>
<td>0.8 to 1.5</td>
<td>0.7–1.7</td>
<td>0.3</td>
</tr>
</tbody>
</table>
Table 4 presents results by study type. For CHC solvents and related compounds, cohort studies with internal reference and case-control studies yielded an MRR of 1.4 (95% CI 0.8 to 2.4; four populations). In studies that used general populations as the reference (SMR/SIR studies), MRR was 1.3 (0.9 to 2.0; 16 populations). For the lower quality proportional studies, it was 1.1 (0.9 to 1.3; four populations).

For nickel and nickel compounds the MRR was highest for case-control studies (2.0; 95% CI 1.2 to 3.2; two populations). For chromium and chromium compounds it was highest in SMR/SIR studies (2.3; 0.9 to 5.8; six populations), as for PAHs it was (3.0; 0.7 to 13.2; two populations).

For asbestos, the 20 SMR/SIR populations yielded a significant MRR of 1.2 (95% CI 1.0 to 1.5), whereas the four case-control populations resulted in an MRR of 0.7 (0.5 to 1.0). Two cohort studies with an internal reference yielded a significant MRR of 1.4 (95% CI 1.3 to 1.9) for diesel engine exhaust, which was, however, not confirmed in case-control and SMR/SIR studies, the overall MRR remaining at 1.0 (95% CI 0.9 to 1.2).

For insecticides, the overall MRR was 1.5 (95% CI 0.6 to 3.7). For the two case-control studies, both based on cytological diagnoses, it was 3.7 (95% CI 0.3 to 43.3), which was not confirmed in the one occupational cohort mortality study. All insecticide results

Table 4 presents results by study type. For CHC solvents and related compounds, cohort studies with internal reference and case-control studies yielded an MRR of 1.4 (95% CI 0.8 to 2.4; four populations). In studies that used general populations as the reference (SMR/SIR studies), MRR was 1.3 (0.9 to 2.0; 16 populations). For the lower quality proportional studies, it was 1.1 (0.9 to 1.3; four populations).

For nickel and nickel compounds the MRR was highest for case-control studies (2.0; 95% CI 1.2 to 3.2; two populations). For chromium and chromium compounds it was highest in SMR/SIR studies (2.3; 0.9 to 5.8; six populations), as for PAHs it was (3.0; 0.7 to 13.2; two populations).

For asbestos, the 20 SMR/SIR populations yielded a significant MRR of 1.2 (95% CI 1.0 to 1.5), whereas the four case-control populations resulted in an MRR of 0.7 (0.5 to 1.0). Two cohort studies with an internal reference yielded a significant MRR of 1.4 (95% CI 1.3 to 1.9) for diesel engine exhaust, which was, however, not confirmed in case-control and SMR/SIR studies, the overall MRR remaining at 1.0 (95% CI 0.9 to 1.2).

For insecticides, the overall MRR was 1.5 (95% CI 0.6 to 3.7). For the two case-control studies, both based on cytological diagnoses, it was 3.7 (95% CI 0.3 to 43.3), which was not confirmed in the one occupational cohort mortality study. All insecticide results

Table 4 presents results by study type. For CHC solvents and related compounds, cohort studies with internal reference and case-control studies yielded an MRR of 1.4 (95% CI 0.8 to 2.4; four populations). In studies that used general populations as the reference (SMR/SIR studies), MRR was 1.3 (0.9 to 2.0; 16 populations). For the lower quality proportional studies, it was 1.1 (0.9 to 1.3; four populations).

For nickel and nickel compounds the MRR was highest for case-control studies (2.0; 95% CI 1.2 to 3.2; two populations). For chromium and chromium compounds it was highest in SMR/SIR studies (2.3; 0.9 to 5.8; six populations), as for PAHs it was (3.0; 0.7 to 13.2; two populations).

For asbestos, the 20 SMR/SIR populations yielded a significant MRR of 1.2 (95% CI 1.0 to 1.5), whereas the four case-control populations resulted in an MRR of 0.7 (0.5 to 1.0). Two cohort studies with an internal reference yielded a significant MRR of 1.4 (95% CI 1.3 to 1.9) for diesel engine exhaust, which was, however, not confirmed in case-control and SMR/SIR studies, the overall MRR remaining at 1.0 (95% CI 0.9 to 1.2).

For insecticides, the overall MRR was 1.5 (95% CI 0.6 to 3.7). For the two case-control studies, both based on cytological diagnoses, it was 3.7 (95% CI 0.3 to 43.3), which was not confirmed in the one occupational cohort mortality study. All insecticide results

Table 4 presents results by study type. For CHC solvents and related compounds, cohort studies with internal reference and case-control studies yielded an MRR of 1.4 (95% CI 0.8 to 2.4; four populations). In studies that used general populations as the reference (SMR/SIR studies), MRR was 1.3 (0.9 to 2.0; 16 populations). For the lower quality proportional studies, it was 1.1 (0.9 to 1.3; four populations).

For nickel and nickel compounds the MRR was highest for case-control studies (2.0; 95% CI 1.2 to 3.2; two populations). For chromium and chromium compounds it was highest in SMR/SIR studies (2.3; 0.9 to 5.8; six populations), as for PAHs it was (3.0; 0.7 to 13.2; two populations).

For asbestos, the 20 SMR/SIR populations yielded a significant MRR of 1.2 (95% CI 1.0 to 1.5), whereas the four case-control populations resulted in an MRR of 0.7 (0.5 to 1.0). Two cohort studies with an internal reference yielded a significant MRR of 1.4 (95% CI 1.3 to 1.9) for diesel engine exhaust, which was, however, not confirmed in case-control and SMR/SIR studies, the overall MRR remaining at 1.0 (95% CI 0.9 to 1.2).

For insecticides, the overall MRR was 1.5 (95% CI 0.6 to 3.7). For the two case-control studies, both based on cytological diagnoses, it was 3.7 (95% CI 0.3 to 43.3), which was not confirmed in the one occupational cohort mortality study. All insecticide results

Table 4 presents results by study type. For CHC solvents and related compounds, cohort studies with internal reference and case-control studies yielded an MRR of 1.4 (95% CI 0.8 to 2.4; four populations). In studies that used general populations as the reference (SMR/SIR studies), MRR was 1.3 (0.9 to 2.0; 16 populations). For the lower quality proportional studies, it was 1.1 (0.9 to 1.3; four populations).

For nickel and nickel compounds the MRR was highest for case-control studies (2.0; 95% CI 1.2 to 3.2; two populations). For chromium and chromium compounds it was highest in SMR/SIR studies (2.3; 0.9 to 5.8; six populations), as for PAHs it was (3.0; 0.7 to 13.2; two populations).

For asbestos, the 20 SMR/SIR populations yielded a significant MRR of 1.2 (95% CI 1.0 to 1.5), whereas the four case-control populations resulted in an MRR of 0.7 (0.5 to 1.0). Two cohort studies with an internal reference yielded a significant MRR of 1.4 (95% CI 1.3 to 1.9) for diesel engine exhaust, which was, however, not confirmed in case-control and SMR/SIR studies, the overall MRR remaining at 1.0 (95% CI 0.9 to 1.2).

For insecticides, the overall MRR was 1.5 (95% CI 0.6 to 3.7). For the two case-control studies, both based on cytological diagnoses, it was 3.7 (95% CI 0.3 to 43.3), which was not confirmed in the one occupational cohort mortality study. All insecticide results
Only agents with both MRRs >1.0 considered. Negative lower confidence bounds of PEFs and EEFs forced to zero.

Table 5  Populations (n), population aetiological fractions (PEFs), aetiological fractions among exposed (EEFs), (95% CIs)

<table>
<thead>
<tr>
<th>Agent</th>
<th>Proportion exposed</th>
<th>Population aetiological fraction (PEF)</th>
<th>Aetiological fraction among exposed (EEF)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>MRR 95% CI</td>
<td>PEF 95% CI</td>
<td>EEF 95% CI</td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Aliphatic and alicyclic hydrocarbon solvents and related compounds</td>
<td>2 0.094 1.3 0.8 to 2.0 0.022 0.00 to 0.047</td>
<td>2 1.3 0.8 to 2.0 0.23 0.00 to 0.50</td>
<td></td>
</tr>
<tr>
<td>Chlorinated hydrocarbon solvents and related compounds</td>
<td>3 0.061 1.2 0.7 to 1.9 0.010 0.00 to 0.029</td>
<td>20 1.4 1.0 to 1.8 0.29 0.00 to 0.44</td>
<td></td>
</tr>
<tr>
<td>Chromium and chromium compounds</td>
<td>3 0.034 1.1 0.7 to 1.6 0.003 0.00 to 0.013</td>
<td>9 1.4 0.9 to 2.3 0.29 0.00 to 0.57</td>
<td></td>
</tr>
<tr>
<td>Fungicides</td>
<td>1 0.003 1.4 0.3 to 7.2 0.001 0.00 to 0.0029</td>
<td>2 1.3 0.4 to 3.8 0.23 0.00 to 0.74</td>
<td></td>
</tr>
<tr>
<td>Insecticides</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Iron and iron compounds</td>
<td>1 0.067 1.3 0.7 to 2.5 0.016 0.00 to 0.040</td>
<td>1 1.3 0.7 to 2.5 0.23 0.00 to 0.60</td>
<td></td>
</tr>
<tr>
<td>Lead and lead compounds</td>
<td>2 0.011 1.1 0.8 to 1.6 0.010 0.00 to 0.041</td>
<td>4 1.1 0.8 to 1.5 0.09 0.00 to 0.33</td>
<td></td>
</tr>
<tr>
<td>Nickel and nickel compounds</td>
<td>2 0.063 2.0 1.2 to 3.2 0.032 0.011 to 0.043</td>
<td>4 1.9 1.2 to 3.2 0.47 0.17 to 0.69</td>
<td></td>
</tr>
<tr>
<td>Polycyclic aromatic hydrocarbons (PAHs)</td>
<td>2 0.029 1.4 0.8 to 2.3 0.008 0.00 to 0.016</td>
<td>4 1.5 0.9 to 2.4 0.33 0.00 to 0.58</td>
<td></td>
</tr>
<tr>
<td>Silica dust</td>
<td>2 0.063 1.5 0.8 to 2.7 0.021 0.00 to 0.040</td>
<td>3 1.4 0.9 to 2.0 0.29 0.00 to 0.50</td>
<td></td>
</tr>
<tr>
<td>Wood dust</td>
<td>2 0.062 1.2 0.9 to 1.6 0.01 0.00 to 0.023</td>
<td>5 1.2 0.9 to 1.6 0.17 0.00 to 0.38</td>
<td></td>
</tr>
</tbody>
</table>

Only agents with both MRRs >1.0 considered. Negative lower confidence bounds of PEFs and EEFs forced to zero. MRR=meta risk ratio, as estimated from population based and hospital based case-control studies for PEF, and all but proportional studies for EEF.

Concerned organochlorine compounds. There were 12 further studies that considered unspecified pesticides. This group of agents was considered to be too heterogeneous in exposures and was dropped from the analysis.

Given the low proportions of the populations that were exposed, the PEFs remained low, from 0.1%–3% (table 5). Assuming independence between exposures, summing up of the PEFs resulted in an overall aetiological fraction of 12% for workplace exposures. The agent specific EEFs ranged from 9% to 47%, with wide 95% CIs. In a subpopulation exposed to nickel and nickel compounds, the EEF was 47% (95% CI 17 to 69%); to PAHs, 33% (0 to 58%); to insecticides, 33% (0 to 73%); to CHC solvents and related compounds, 29% (0 to 44%); to chromium and chromium compounds, 29% (0 to 57%); and to aliphatic and alicyclic solvents, 23% (0 to 50%).

Discussion
EVIDENCE

The excess risk found for CHC solvents and related compounds was based on 20 populations. Heterogeneity of RRs was nearly significant and may be explained by differences in the quality and exposure level of the agents. Various compounds with variable carcinogenic potential were mentioned as worker exposures: trichloroethylene, tetrachloroethylene, 1,1,1-trichloroethane, methane chloride, vinyl chloride, ethylene chlorohydrine, ethylene dichloride, bis(chloromethyl)ether, and polychlorinated biphenyls. Intensities and long term doses were characterised in most of the studies either poorly or not at all.

The risk for nickel and nickel compounds was most evident in population based case-control studies. For chromium and chromium compounds, the MRR was non-significantly increased in all studies, but was not in excess in population based case-control studies. For PAHs, a non-significant increase was present in all studies, in population-based case-control studies, and in the two SMR/SIR studies. These findings could have occurred by chance.

The excess of silica dust reached significance in one of three studies. Finnish population-based case-control study found a significant excess for aliphatic and alicyclic hydrocarbon solvents. This finding was aggregated with the finding of no excess for alkanes (C5-C17) in another population based study from Montreal, the result being an MRR of 1.3 (95% CI 0.8 to 2.0).

Two case-control studies, both based on cytological diagnoses, one SMR study, and one PMR study considered exposure to organochlorine insecticides. The insecticidal agents listed as exposures were 1,1-bis(4-chlorophenyl)-2,2,2-trichloroethane (DDT), chlor dane, heptachlor, endrin, aldrin, dieldrin, bulan, chlorfenethol, chloropropylate, dicofol, ethylan, methoxychlor, and tetrachlorodiphenylethane (TDE). Excluding the PMR study, the aggregated MRR was 1.5 (95% CI 0.6 to 3.7). In case-control studies it was 3.7 (0.3 to 43) based on the random effects model, and 1.7 (0.9 to 3.2) based on the fixed effects model (heterogeneity p=0.09). The highest RR was obtained for exposure to the DDT family (DDT, ethylan, DDD; OR 21.0; 95% CI 2.6 to 96; five exposed cases) in a case-control study nested in a chemical manufacturing cohort.

Potential confounders included nitrophenol derivatives, clays, N,N-dimethylformamide, dispersing agents, octane, and carbon tetrachloride. The other case-control study was population based (Michigan, US), with self reported exposures. Based on 21 exposed cases, it yielded an OR of 1.5 (95% CI 0.8 to 2.9) for organochlorine insecticides. Assuming an effect, the difference between the two point estimates might be due to qualitative and quantitative differences in exposures between manufacturing and agricultural application.

There is a possibility of effect modification of environmental or occupational determinants by lifestyles (tobacco, alcohol, coffee) or dietary factors. These interactions were not considered in the studies. Also, genetic factors may interact with environmental or occupational exposures.

COMBINABILITY AND HETEROGENEITY

Epidemiological meta-analyses have imperfect combinability of results associated with different study types, methods, populations, exposure circumstances, and diagnostic specificities. We calculated MRRs including and excluding the poorest quality studies (proportional studies). We also calculated separate MRRs for cohort studies with internal controls,
Occupational exposures and pancreatic cancer

Several populations were poorly characterised. There were even studies that did not specify whether the cohort consisted of men, women, or both. We analysed the data for known male and female populations separately, and found data for women to be associated with a slightly higher MRR than data for men for CHC solvents and related compounds. There was in all likelihood substantial heterogeneity across populations in the quality and intensity of exposure categories, in the intake route (respiratory, dermal, or other) of exposure, time aspects of exposure (period, latency, duration, quality, and intensity), and applied scales of exposure, as well as in the quality of diagnosis of pancreatic cancer. Qualitative and quantitative differences in exposures have already been exemplified in connection with CHC solvents and related compounds and organochlorine insecticides. Based on a rough statistical test, significant heterogeneity of risk was found for asbestos, electromagnetic fields, CHC solvents and related compounds, and man made vitreous fibres. Some studies did not document the exposure aspects at all, and no study provided a full documentation. Expert assessments, which represent an imperfect yet acceptable method of exposure assessment, were used in 25 populations. Industrial hygiene measurements that represent a certain degree of objectivity were used as the prime source of exposure data in only four populations. Exposure assessment based on job titles (57 populations) are of lower quality, unless the exposures are highly homogeneous within job titles. However, some of the populations represented rather homogeneous single title cohorts. Job exposure matrices assess exposures better if the matrix is specific for branch and job title, even for company and period. The JEMs of variable degree of specificity were applied in 15 populations. Most were relatively unspecific and thereby induced exposure misclassification. Misclassification, however, was likely to be non-differential, resulting in underestimation of the MRR. Multiple sources of exposure data were applied in 37 populations. Exposure data were longitudinal in 127 populations and lifelong in 47 populations. The longitudinality of exposures was thus well covered.

Misclassification rates for pancreatic cancer are marked, particularly when the diagnosis is based on death certificate only. This adds to the bias in the meta estimates towards the null value. Garabrantd et al[2] showed what the magnitude of the impact may be in a single study. They compared, in a case-control study of pancreatic cancer, ORs for the DDT family between cases representing death certificates and cases representing cytological verification. For death certificate cases the ORs ranged from 0.8 to 2.6; for cytologically verified cases, from 15.4 to infinity. In our meta-analysis, agent specific MRRs were higher for nine agents but lower for 10 agents in populations in which histological verification was applied, compared with no histological verification. For CHC solvents and related compounds, the two MRRs were practically identical. For PAHs and chromium, the MRRs were higher in populations with histological diagnoses.

PUBLICATION BIAS
Publication bias is not likely in this study, as very few small studies expressly considered the occupational determinants of pancreatic cancer. Non-positive occupational findings from small studies therefore were not likely to remain unpublished. Also, a funnel plot for CHC solvents and related compounds, for which an excess risk was found (figure), did not identify a concentration of small studies (high SE of ln (RR)) at high RRs. Publication bias therefore may be claimed to be minimal or non-existent. Negative results based on reasonably large numbers are unlikely to remain unpublished.

A counterargument may be raised about cohort studies with multiple end points. Some of these studies deleted results based on small numbers, occasionally for pancreatic cancer. This omission may have minor influence on the metaresults. Some case-control studies may have omitted results for rare exposures, with similar minor effect on the metaresults.

SELECTION OF STUDIES
We used major databases and lists of references of the studies for the identification of studies. Studies not found in major databases are probably of lower quality. For the same reason and because of the cost we did not try to identify unpublished studies.

EXTRACTION
Extractor bias was minimised by the formal extraction procedure between the extractors, and the central checking of the extraction. The procedure was also intended to guarantee the extraction of the relevant risk estimates in studies that offered several alternative risk estimates.
REFERENCE POPULATIONS
Not all populations in the data were strictly independent because in some studies an internal unexposed industrial population was used as the common reference for more than one exposed population. This was rare, however, and we consider the ensuing bias in the precision of MRR to be minimal. Reference populations are a problem in proportional studies, where the population basis is unknown. For most analyses, we excluded proportional studies for this reason. Comparability of populations may be a problem in SMR and SIR studies because of the healthy worker effect. It is unknown to what extent the healthy worker effect and its components might have biased our metaregressions.

CONFOUNDING
Control of confounding is a problem in studies of pancreatic cancer, as tobacco smoking and diabetes are the only known common causes of this malignancy. Even a rough measurement of confounding bias is difficult. Case-control studies are in principle best equipped for adjustment for confounders. Several of them did adjust for various factors. In other study types, adjustment was rare. Attempts to aggregate results over studies that adjusted for smoking failed because of small numbers of such studies. The number of available populations for which adjustment was done ranged from zero to two across the occupational agents.

AETIOLOGICAL FRACTIONS
Aetiological fractions were unstable both statistically and substantially. The PEFs were highly dependent of the variable occupational exposure patterns across populations and periods. The variability concerns the proportions of exposed subjects, and intensities and time patterns of exposures. Also, the technical definitions of exposed varied between studies. The same holds for EEFs, with the exception that the proportion of exposed subjects is not a concern.

Conclusions
The aetiological fraction of pancreatic cancer due to occupational exposures within a population was estimated at 12%. This may be an underestimate because of misclassification of exposures and end points in the studies. The aetiological fractions among exposed subjects was highest at 47% in people occupationally exposed population. This was rare, however, as the common reference for more than one exposed population. The excess may be pronounced in women. Excesses associated with occupational exposure to chromium and chromium compounds and nickel and nickel compounds were suggested. More limited evidence was found for organochlorine insecticides, silica dust, and aliphatic and alicyclic hydrocarbon solvents.

Future research into the aetiology of pancreatic cancer should concentrate on more refined assessment of exposure and end points, assessment of interactions between occupational and environmental factors, lifestyle (tobacco, alcohol, coffee, diet), and interactions between genes and the environment. Future research therefore calls for studies, complex measures, and refined statistical methods.

The Finnish Work Environment Fund has supported this study. Ms Riitta Järnström and Mr Tuomas Hanhinen created the reference database.

Occupational exposures and pancreatic cancer

24 Brown DP. Mortality of workers employed at organochlo-
rine pesticide manufacturing plants: an update. Scand J

25 Brown DP, Dement JM, Okan A. Mortality patterns among
female and male chrysolite asbestos textile workers. J

follow up of cancer mortality among workers in a thero-
electric power plant. Scand J Work Environ Health 1986;12:
631–2.

27 Checkoway H, Heyer NJ, Seixas NS, et al. Dose-response
association of silica with non-malignant respiratory
disease and lung cancer mortality in the diatomaceous

28 Checkoway H, Ference LD. Mortality among PVC-fabricating

29 Checkoway H, Frenzel-Beyme R. Mortality of workers in a
German rock-wool factory: a second look with extended follow

30 Checkoway H, Hsu P, Boffetta P, Morabito of Italian
lead smelter workers. Scand J Work Enviro Health 1997;23:15–
23.

of the occupational exposure to DD T. A preliminary report.

32 Coggon D, Pannett B, Winter PD, et al. Mortality among chemi-
ical asbestos textile workers: cohort mortality and

658.

34 Decouflé P. Further analysis of cancer mortality patterns
among workers exposed to cutting oil mists. J Natl Cancer

35 Decouflé P, Blattner WA, Blair A. Mortality among chemi-

36 Dement JM, Brown DP, Okan A. Follow up study of chrys-
olite asbestos textile workers: cohort mortality and

37 Enterline PE, Hartley J, Henderson V. Asbestos and cancer:
73.

38 Enterline PE, Day R, Marsh GM. Cancers related to expo-
sure to 2 methyl-4 chlorophenoxyacetic acid. Scand J

39 Franchini I, Magnani F, Mutti A. Mortality experience among
chromatographing workers. Initial findings. Scand J

40 Frentzel-Beyme R, Thiess AM, Wieland R. Survey of mor-
talities in workers exposed to cutting oil mists: an
occupational cohort study. Occup Environ Med 1990;47:
1025–30.

41 Fryzek JP, Garabrant DH, Harlow SD, et al. A case-control
study of self reported exposures to pesticides and pancre-
cas cancer in southeastern Michigan. Int J Cancer 1997;72:
67–7.

42 Garabrant DH, Held J, Homa D. Response (DDT and pan-

workers exposed to formaldehyde in the British chemical

44 Gibbs GW, Ansell J, Soden K. A cohort mortality study of
cellulose fibre production workers exposed to methylene

study of workers exposed to toluene in the plastic

46 Gustavsson P, Reuterwall C. Mortality and incidence of
169–74.

47 Haukka J, Olsen JH. Formaldehyde and cancer morbidity
among male employees in Denmark. Cancer Causes Control

48 Hearn CD, Pfifer JW, Grose F. Absence of adverse mortality
among ferrochromium and ferrosilicon workers: an ex-


50 Hansen J, Olsen JH. Formaldehyde and cancer morbidity
among male employees in Denmark. Cancer Causes Control


52 Heelan DK, Jansen A, Schraub CM, et al. Excess mortality
among workers exposed to formaldehyde in the garment

among workers exposed to formaldehyde in the garment

54 Heelan DK, Jansen A, Schraub CM, et al. Excess mortality
among workers exposed to formaldehyde in the garment

among workers exposed to formaldehyde in the garment

among workers exposed to formaldehyde in the garment

---

**Occupational and Environmental Medicine** - http://www.occenvmed.com

Visitors to the world wide web can now access Occupational and Environmental Medicine either through the BMJ Publishing Group’s home page (http://www.bmjng.com) or directly by using its individual URL (http://www.occenvmed.com). There they will find the following:

- Current contents list for the journal
- Contents lists of previous issues
- Members of the editorial board
- Subscribers’ information
- Instructions for authors
- Details of reprint services.

A hotlink gives access to:

- BMJ Publishing Group home page
- British Medical Association website
- Online books catalogue
- BMJ Publishing Group books.

The web site is at a preliminary stage and there are plans to develop it into a more sophisticated site. Suggestions from visitors about features they would like to see are welcomed. They can be left via the opening page of the BMJ Publishing Group site or, alternatively, via the journal page, through “about this site”.

---

Downloaded from [http://oem.bmj.com/](http://oem.bmj.com/) on June 18, 2017 - Published by [group.bmj.com](http://group.bmj.com)
Occupational exposures and pancreatic cancer: a meta-analysis

I Anneli Ojajärvi, Timo J Partanen, Anders Ahlbom, Paolo Boffetta, Timo Hakulinen, Nadia Jourenkova, Timo P Kauppinen, Manolis Kogevinas, Miquel Porta, Harri U Vainio, Elisabete Weiderpass and Catharina H Wesseling

*Occup Environ Med* 2000 57: 316-324
doi: 10.1136/oem.57.5.316

Updated information and services can be found at:
http://oem.bmj.com/content/57/5/316

These include:

**References**

This article cites 91 articles, 16 of which you can access for free at:
http://oem.bmj.com/content/57/5/316#BIBL

**Email alerting service**

Receive free email alerts when new articles cite this article. Sign up in the box at the top right corner of the online article.

**Topic Collections**

Articles on similar topics can be found in the following collections

Other exposures (1023)

**Notes**

To request permissions go to:
http://group.bmj.com/group/rights-licensing/permissions

To order reprints go to:
http://journals.bmj.com/cgi/reprintform

To subscribe to BMJ go to:
http://group.bmj.com/subscribe/